

# Isolation and characterization of a heterotrophic nitrifier *Proteus mirabilis* strain V7 and its potential application in $\text{NH}_4^+$ -N removal

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**Abstract** Ammonia (referring to both  $\text{NH}_3$  and  $\text{NH}_4^+$ ), is one of the most toxic nitrogen forms, and is toxic to many aquatic organisms. The present study identified a heterotrophic nitrifier *Proteus mirabilis* strain V7, isolated from the coastal seawater. Almost 100 % of  $82 \text{ mg L}^{-1}$   $\text{NH}_4^+$ -N was removed within 48 h, while 51.2 % of total inorganic nitrogen was removed within 96 h after inoculation of *Proteus mirabilis* strain V7. Moreover, *Proteus mirabilis* strain V7 could remove  $\text{NH}_4^+$ -N in a seawater matrix, but with much lower efficiency. Three water samples containing  $\text{NH}_4^+$ -N at

concentrations of 64.40, 18.00 and  $9.82 \text{ mg L}^{-1}$  were chosen to further test the  $\text{NH}_4^+$ -N removal ability of *Proteus mirabilis* strain V7. All of the  $\text{NH}_4^+$ -N was removed, with only a trace production of  $\text{NO}_2^-$  and  $\text{NO}_3^-$ . A putative subunit of ammonia monooxygenase from *Proteus mirabilis* strain V7,  $\text{amoA}_{\text{Pm}}$ , was cloned and its mRNA level in the presence of  $\text{NH}_4^+$  was characterized by real time reverse transcriptase PCR (RT-PCR). A 1.8-fold increase in the mRNA level of  $\text{amoA}_{\text{Pm}}$  was observed, which indicated that AmoA may be involved in  $\text{NH}_4^+$  oxidization. Our results indicated that *Proteus mirabilis* strain V7 could potentially be used as new resource for the treatment of  $\text{NH}_4^+$ -N contaminated environments.

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## Introduction

Ammonia in the form of un-ionized ammonia ( $\text{NH}_3$ ) or ionized ammonia ( $\text{NH}_4^+$ ) is an inorganic and dissolved nitrogen compound that can be detected frequently in the aquatic environment, especially where dissolved oxygen is lacking (Alcaraz et al. 1999). In solution, the acid ionization of  $\text{NH}_3$  is realized via the equation described by Whitfield (1974). The relative proportion of  $\text{NH}_4^+$  and  $\text{NH}_3$  depends mainly on environmental factors such as temperature, salinity and pH of the aqueous solution (Alcaraz et al. 1999). As an important nitrogen compound, ammonia within the range of  $2\text{--}10 \text{ mg L}^{-1}$  has been proved to be severely toxic to most aquatic organisms. The acceptable ammonia concentration in drinking water has been designated as  $1.5 \text{ mg L}^{-1}$  by the US Environmental Protection Agency (US EPA) (WHO 1990; US EPA 2009).

Considering the hazardous effects of ammonia, methods to efficiently remove ammonia from aquatic environments have been urgently recommended (US EPA 1993; Taylor et al.

2009). Among the methods developed, biological transformation through the metabolism of nitrifying bacteria is a superior approach due to its advantages of higher efficiency and cost-effectiveness over the alternative physical and chemical processes such as ion exchange and adsorption (US EPA 1993; Rostron et al. 2001; Jorgensen and Weatherley 2003; Khardenavis et al. 2007; Andrade do Canto et al. 2008). Traditionally,  $\text{NH}_4^+$  was subsequently oxidized to  $\text{NO}_3^-$  by two kinds of ubiquitous chemoautotrophic nitrifying bacteria, ammonium oxidizing bacteria (AOB) and nitrite oxidizing bacteria (NOB), via the intermediate of  $\text{NO}_2^-$  (Ward 1996; Schaechter 2009). Recently, studies on the  $\text{NH}_4^+$ -N removal depending on  $\text{NH}_4^+$  oxidization by heterotrophic nitrifying bacteria, which also possess the ability to oxidize  $\text{NH}_4^+$  but catabolize organic compounds in a life mode different from that of traditional chemoautotrophic nitrifying bacteria, are increasing (Winogradsky 1890; Witzel and Overbeck 1979; Papen and von Berg 1998; Taylor et al. 2009; Zhao et al. 2010; Yang et al. 2011; Sarioglu et al. 2012). Compared with traditional chemoautotrophic nitrifying bacteria, heterotrophic nitrifying bacteria not only possess relatively high growth rate and nutrient utilization efficiency but also have the advantage of higher tolerance to  $\text{NH}_4^+$  and organic matter (Zhao et al. 2010). Many heterotrophic nitrifying bacteria, including *Alcaligenes faecalis*, *Acinetobacter calcoaceticus*, *Thiosphaera pantotropa*, *Bacillus* sp., *Microvirgula aerodenitrificans*, *Pseudomonas stutzeri*, *Pseudomonas putida*, *Comamonas* sp. and *Diaphorobacter* sp., have been characterized for their potential application in bioremediation of  $\text{NH}_4^+$  or other nitrogenous pollutants (Otte et al. 1996; Patureau et al. 1997; Daum et al. 1998; Su et al. 2001; Kim et al. 2005; Khardenavis et al. 2007; Taylor et al. 2009; Zhao et al. 2010; Yang et al. 2011).

In the present study, a *Proteus mirabilis* strain V7 isolated from a coastal seawater sample was newly identified as a heterotrophic nitrifying bacterium. The products of  $\text{NH}_4^+$  oxidization by this heterotrophic nitrifying bacterium *P. mirabilis* strain V7 were primarily determined. The gene encoding a putative subunit of ammonia monooxygenase (AMO), *amoA*<sub>Pm</sub>, was cloned and its relative mRNA level in the presence and absence of  $\text{NH}_4^+$  was analyzed.

## Materials and methods

### Strains, medium, growth conditions and chemicals

*Escherichia coli* strains Top10 and BL21(DE3) used in this study were grown in Luria-Bertani (LB) medium at 37 °C with appropriate antibiotics. The antibiotics ampicillin (Ap) and kanamycin (Kn) were used at working concentrations of 100 mg L<sup>-1</sup> and 50 mg L<sup>-1</sup>, respectively. Five environmental

bacteria, i.e., *Proteus mirabilis* strain V7, *Bacillus cereus* strain KF8, *Bacillus pumilus* strain Bb, *Candida boidinii* strain Cb and *Pseudomonas putida* strain SP1 isolated from coastal seawater, were grown in specific medium at 30 °C. LB medium consisted of 10 g tryptone, 5 g yeast extract, and 10 g NaCl in 1 L ddH<sub>2</sub>O. Synthetic succinate medium contained 6.0 g K<sub>2</sub>HPO<sub>4</sub>, 3.0 g KH<sub>2</sub>PO<sub>4</sub>, 1.0 g (NH<sub>4</sub>)<sub>2</sub>SO<sub>4</sub>, 0.1 g MgSO<sub>4</sub>·7H<sub>2</sub>O, and 4.0 g succinic acid in 1 L ddH<sub>2</sub>O. Phosphate succinate medium without (NH<sub>4</sub>)<sub>2</sub>SO<sub>4</sub> consisted of 6.0 g K<sub>2</sub>HPO<sub>4</sub>, 3.0 g KH<sub>2</sub>PO<sub>4</sub>, 0.1 g MgSO<sub>4</sub>·7H<sub>2</sub>O, and 4.0 g succinic acid in 1 L ddH<sub>2</sub>O. 2216E medium consisted of 5 g tryptone, 1 g yeast extract, and 0.01 g FePO<sub>4</sub> in 1 L aged seawater.

Plasmid pBS-T was purchased from Transgen Biotech (Beijing, China). Restriction endonucleases, alkaline phosphatase, T4 DNA ligase and *Taq* DNA polymerase were purchased from Fermentas (MBI, Beijing, China). Plasmid preparation, the extraction of DNA fragments from agarose gels, and the purification of PCR products were performed using the respective kits from Omega Bio-Tek (Norcross, GA) according to the manufacturer's instructions. Most of the other chemicals used in this study were purchased from Sangon at analytical grade (Sangon, Shanghai, China).

### Bacteria selection by semi-quantitative ammonia detection

Selection of a bacterium capable of decreasing the concentration of  $\text{NH}_4^+$ -N was carried out using semi-quantitative ammonia detection kits (Luheng Biotech, Hangzhou, China). Briefly, supernatants from bacterial cultures of *Proteus mirabilis* strain V7, *B. cereus* strain KF8, *B. pumilus* strain Bb, *C. boidinii* strain Cb and *Pseudomonas putida* strain SP1 grown in synthetic succinate medium, were diluted ten times into 15 mL tubes. One pack of reagent I was added to each tube, followed by gently shaking and standing at room temperature for 2 min. Then a pack of reagent II was added, followed by mixing uniformly and standing at room temperature for another 15 min. The developed color of each tube was compared with the standard board to determine the remaining  $\text{NH}_4^+$ -N semi-quantitatively.

### DNA techniques

One pair of primers PmamoF1 (5'-GACGCATATGTCATTTTAAAAAACTACC-3') and PmamoR1 (5'-CTCGAGCTCACGTTTTCTTTTCTG-3') was designed based on the nucleotide sequence of AM942759 to amplify the putative *amoA* gene from *Proteus mirabilis* (*amoA*<sub>Pm</sub>). Another pair of primers PmamoF2 (5'-TACACTGACTTTTTATCCAACCTGACCT-3') and PmamoR2 (5'-ACAGCACCTAACATAGCC TCTTCT-3'), was designed for real time reverse transcriptase PCR (RT-PCR) to analyze relative mRNA levels of *amoA*<sub>Pm</sub>. All primers used in this study were synthesized by Sunny

Biotech (Shanghai, China). PCR amplifications were carried out in a Tpersonal thermocycler (Biometra, Goettingen, Germany). PCR products were separated on 0.8 % agarose gel, and DNA fragments were purified and ligated into specific vectors. The ligation mixture was transformed into the competent cells of *E. coli* strains Top10, and the transformants were selected on LB plates supplemented with appropriate antibiotic. DNA sequencing was carried out by Sunny Biotech (Shanghai, China). To obtain the purified AmoA<sub>Pm</sub>, a plasmid pETamoA<sub>Pm</sub> was constructed, in which a PCR product amplified with primers PmamoF1 and PmamoR1, was digested simultaneously with *Nde*I and *Xho*I, and then inserted into pET258, constructed by Zhang and Sun (2007), between *Nde*I/*Xho*I sites.

### Real time RT-PCR

Real time RT-PCR was carried out according to the description of Zhang et al. (2008). Briefly, total RNA was extracted from cells grown in the presence and absence of NH<sub>4</sub><sup>+</sup>. Real time RT-PCR was carried out in an ABI 7500 real-time detection system using a Sybr ExScript RT-PCR kit (Takara Biotech, Dalian, China). The primers of *amoA*<sub>Pm</sub> used for RT-PCR were PmamoF2 and PmamoR2. 16S rRNA was used as the reference gene. The primers of 16S rRNA used for the RT-PCR were 933 F (5'-GCACAAGCGGTGGAGCATGTGG-3') and 16SRTR1 (5'-CGTGTGTAGCCCTGGTCGTA-3'). Dissociation analysis of amplification products was performed at the end of each PCR to confirm that only one PCR product was amplified. The comparative threshold cycle method ( $2^{-\Delta\Delta CT}$  method) was used to analyze the relative mRNA level of *amoA*<sub>Pm</sub>.

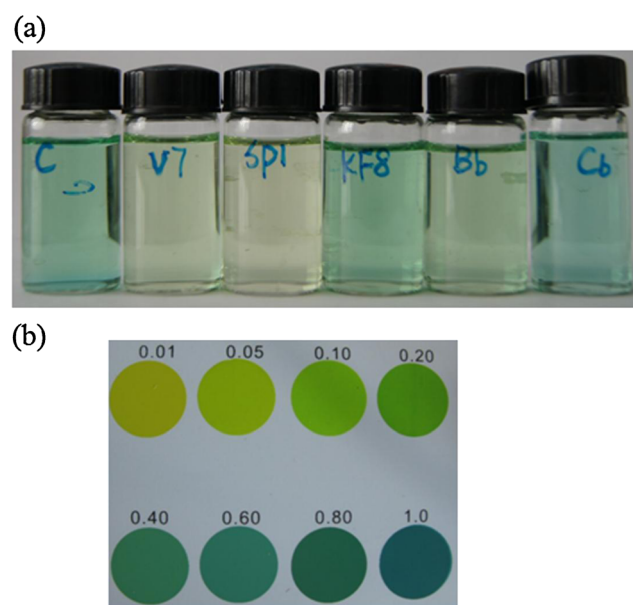
### Expression of recombinant AmoA<sub>Pm</sub>

*Escherichia coli* strain BL21(DE3) harboring the plasmid pETamoA<sub>Pm</sub> was cultured to an OD<sub>600</sub> of 0.5, and expression of AmoA<sub>Pm</sub> was expected to be induced by 1 mM isopropyl-β-D-thiogalactopyranoside (IPTG). After grown at 30 °C for additional 6 h, bacterial cells were harvested by centrifugation and the cell pellet was resuspended in protein denaturing buffer consisted of 100 mM NaH<sub>2</sub>PO<sub>4</sub>, 10 mM Tris-Cl and 8 M urea. Then the cell lysate was centrifuged at 13,000 g for 20 min, and supernatant containing the whole cell proteins was applied for SDS-PAGE analysis.

### NH<sub>4</sub><sup>+</sup>-N removal by *Proteus mirabilis* strain V7

To investigate the effect of initial NH<sub>4</sub><sup>+</sup>-N concentration on the removal efficiency of *Proteus mirabilis* strain V7, cells were inoculated into synthetic succinate medium containing initial NH<sub>4</sub><sup>+</sup>-N concentrations of 20, 50, 100, 200, 500, 1000 and 2000 mg L<sup>-1</sup>. After culturing at 30 °C for 64 h,

supernatants were collected and analyzed for the remaining NH<sub>4</sub><sup>+</sup>-N. The supernatant from the medium containing the same concentration of NH<sub>4</sub><sup>+</sup>-N without inoculation of bacterial cell was used as a control. To investigate whether *Proteus mirabilis* strain V7 could remove NH<sub>4</sub><sup>+</sup>-N from a seawater matrix, 1.0×10<sup>7</sup> CFU mL<sup>-1</sup> bacterial cells were inoculated into 2216E medium containing initial concentrations of NH<sub>4</sub><sup>+</sup> at 20, 50, 100, 200, 500, 1000 and 2000 mg L<sup>-1</sup>, and the remaining NH<sub>4</sub><sup>+</sup>-N in the supernatant was measured at different time points. To determine the products of NH<sub>4</sub><sup>+</sup> oxidization by *Proteus mirabilis* strain V7, experiments were also conducted in a 1-L flask inoculated with 1.0×10<sup>7</sup> CFU mL<sup>-1</sup> bacterial cells in synthetic succinate medium containing 80 mg L<sup>-1</sup> NH<sub>4</sub><sup>+</sup>-N. Aliquots were collected at different time points and supernatants were analyzed for the concentrations of NH<sub>4</sub><sup>+</sup>-N, NO<sub>2</sub><sup>-</sup>-N and NO<sub>3</sub><sup>-</sup>-N. Several water samples were collected around the mouths of several rivers around Yantai, and three samples containing the highest levels of NH<sub>4</sub><sup>+</sup>-N were chosen as the matrices for NH<sub>4</sub><sup>+</sup>-N removal by *Proteus mirabilis* strain V7. After incubated at 30 °C for 48 h, the supernatants were collected and analyzed for concentrations of NH<sub>4</sub><sup>+</sup>-N, NO<sub>2</sub><sup>-</sup>-N and NO<sub>3</sub><sup>-</sup>-N. Measurement of cell density and concentrations of NH<sub>4</sub><sup>+</sup>-N, NO<sub>2</sub><sup>-</sup>-N, and NO<sub>3</sub><sup>-</sup>-N was carried out as follows. The OD<sub>600</sub> of the bacterial culture was measured with a spectrophotometer (Thermo NanoDrop, <http://www.nanodrop.com>). NH<sub>4</sub><sup>+</sup>-N was determined by Nessler's reagent photometry (APHA



**Fig. 1** Selection of bacterial strains capable of removing NH<sub>4</sub><sup>+</sup>-N using a semi-quantitative ammonia nitrogen detection kit. The five strains used were *Proteus mirabilis* strain V7, *Pseudomonas putida* strain SP1, *Bacillus cereus* strain KF8, *B. pumilus* strain Bb and *Candida boidinii* strain Cb. **a** Supernatants from bacterial cultures were reacted with reagents I and II to develop the color. **b** The standard color board used for semi-quantitative detection of NH<sub>4</sub><sup>+</sup>-N; units of concentration on the board in mg L<sup>-1</sup>

**Table 1** Conversion of nitrogen by consumption of  $\text{NH}_4^+$  by *Proteus mirabilis* strain V7. TIN Total inorganic nitrogen

Different nitrogen form	Concentrations of each nitrogen form at different time points					
	0 h	12 h	24 h	48 h	72 h	96 h
$\text{NH}_4^+\text{-N}^a$	81.97±6.3	54.62±4.9	10.68±0.9	0	0	0
$\text{NO}_2^-\text{-N}^b$	2.55±0.06	5.60±0.5	10.66±0.4	13.16±0.3	16.32±0.2	33.11±1.7
$\text{NO}_3^-\text{-N}^a$	0	26.94±1.1	70.34±6.9	79.47±2.7	73.811±7.5	39.98±3.4
TIN <sup>a</sup>	81.97	81.56	81.02	79.47	73.81	39.98
Lost $\text{NH}_4^+\text{-N}$	0	27.35	71.29	81.97	81.97	81.97
Lost TIN	0	0.41	0.95	2.50	8.16	41.99

<sup>a</sup>Unit of concentrations of  $\text{NH}_4^+$ ,  $\text{NO}_3^-$  and TIN =  $\text{mg L}^{-1}$

<sup>b</sup>Unit of concentration of  $\text{NO}_2^-$  =  $\mu\text{g L}^{-1}$

1995).  $\text{NO}_2^-\text{-N}$  was determined by the N-(1-naphthalene)-diaminoethane photometry method.  $\text{NO}_3^-\text{-N}$  was analyzed by the phenol disulfonic acid method.

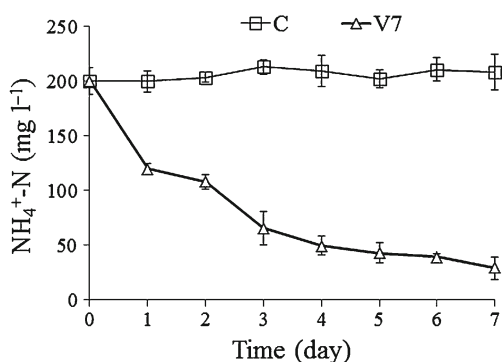
Nucleotide sequence accession number and strain accession numbers

The nucleotide sequence of *amoA*<sub>Pm</sub> has been deposited in the GenBank database with accession number KC775377. The isolate of *Proteus mirabilis* strain V7 was deposited in the China General Microbiological Culture Collection (CGMCC, Beijing, China) with accession number CGMCC 4312.

## Results and discussion

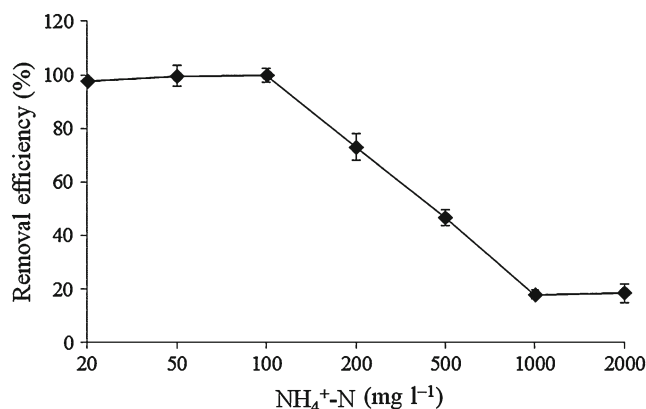
Isolation of a bacterial strain capable of decreasing the concentrations of  $\text{NH}_4^+\text{-N}$

*Proteus mirabilis* strain V7, *B. cereus* strain KF8, *B. pumilus* strain Bb, *C. boidinii* strain Cb, and *Pseudomonas putida* strain SP1 were cultured to detect whether they have the ability to decrease the concentration of  $\text{NH}_4^+\text{-N}$  in the culture medium. *Pseudomonas putida* strain SP1 was used as a positive control according to the study of Daum et al.

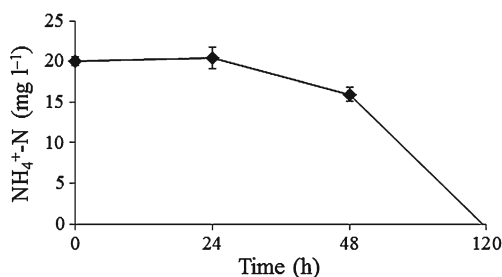


**Fig. 2**  $\text{NH}_4^+\text{-N}$  removal by *Proteus mirabilis* strain V7 over time. The remaining  $\text{NH}_4^+\text{-N}$  in the supernatants from cultures of *Proteus mirabilis* strain V7 was analyzed. The  $\text{NH}_4^+\text{-N}$  in the supernatant from medium without inoculation of bacterial cells was used as a control. Data are the means of three independent experiments and are presented as means±SE

(1998), while none of the other strains has been reported to be able to oxidize  $\text{NH}_4^+$ . After cultured at 30 °C for 5 days, the supernatants were collected and used for the semi-quantitative analysis of the remaining  $\text{NH}_4^+\text{-N}$ . As shown in Fig. 1, the color of supernatants from cultures of *Proteus mirabilis* strain V7 and *Pseudomonas putida* strain SP1 turned yellow-green, which meant that the concentrations of  $\text{NH}_4^+\text{-N}$  in the two supernatants decreased below 4.0  $\text{mg L}^{-1}$ , while the color of supernatants from cultures of *B. cereus* strain KF8, *B. pumilus* strain Bb and *C. boidinii* strain Cb remained the same as, or showed little difference from, the color of the supernatant from the control medium. Although inoculation of *Pseudomonas putida* strain SP1 could also lead to a decrease in the concentration of  $\text{NH}_4^+\text{-N}$ , the growth of *Pseudomonas putida* strain SP1 was poor due to the cell self-lysis. Thus, *Proteus mirabilis* strain V7 was selected for further study on  $\text{NH}_4^+$  oxidization. In the absence of any organic carbon source, *Proteus mirabilis* strain V7 was defective both in cell growth and  $\text{NH}_4^+\text{-N}$  removal activity (data not shown). However, in the presence of succinic acid as the organic carbon source, both growth of *Proteus mirabilis* strain V7 and decreased concentration of  $\text{NH}_4^+\text{-N}$  were observed. According to the results obtained, *Proteus mirabilis* strain



**Fig. 3**  $\text{NH}_4^+\text{-N}$  removal efficiency of *Proteus mirabilis* strain V7 under different initial  $\text{NH}_4^+\text{-N}$  concentrations. The  $\text{NH}_4^+\text{-N}$  remaining after inoculation of *Proteus mirabilis* strain V7 was determined and removal efficiency was defined as the percentage of  $\text{NH}_4^+\text{-N}$  that was initially added to the medium. Data are the means of three independent experiments and are presented as means±SE



**Fig. 4** The  $\text{NH}_4^+$ -N remaining in 2216E medium after inoculation of *Proteus mirabilis* strain V7. The initial concentration of  $\text{NH}_4^+$ -N was  $20 \text{ mg L}^{-1}$ . The remaining  $\text{NH}_4^+$ -N in the supernatant was determined after inoculation at  $30^\circ\text{C}$  for 64 h. Data are the means of three independent experiments and are presented as means $\pm$ SE

V7 was defined as a heterotrophic nitrifying bacterium in the light of the previous references (Alexander 1977; Papen et al. 1989). Moreover, this was the first time that *Proteus mirabilis* strain has been characterized as a heterotrophic nitrifier, and it was expected that *Proteus mirabilis* strain V7 could be potentially applied in  $\text{NH}_4^+$ -N removal.

#### Products of $\text{NH}_4^+$ oxidization by *Proteus mirabilis* strain V7

The concentrations of  $\text{NH}_4^+$ -N,  $\text{NO}_2^-$ -N and  $\text{NO}_3^-$ -N in the supernatants of culture were measured during  $\text{NH}_4^+$  oxidization at different time points. As shown in Table 1, the  $\text{NO}_2^-$ -N concentration in the medium increased to trace levels ( $\mu\text{g L}^{-1}$ ) as the culture time increased, whereas the  $\text{NO}_3^-$ -N concentration increased to  $79.47 \text{ mg L}^{-1}$  as the culture time rose to 48 h. In total, all the  $\text{NH}_4^+$ -N was removed within 48 h, while 51.2 % of the total inorganic nitrogen (TIN) was removed within 96 h of inoculation with cells of *Proteus mirabilis* strain V7. Thus, it could be concluded that *Proteus mirabilis* strain V7 possessed higher  $\text{NH}_4^+$ -N removal efficiency but lower TIN removal efficiency. It is generally accepted that the heterotrophic nitrification and aerobic denitrification can take place simultaneously in one heterotrophic nitrifier through subsequent oxidization of  $\text{NH}_4^+$  to  $\text{NO}_2^-$  and  $\text{NO}_3^-$ , and then reduction of  $\text{NO}_3^-$ , especially  $\text{NO}_2^-$ , to gaseous nitrogen (Castignetti and Hollocher 1982, 1984; Robertson and Gijs Kuenen 1984; Robertson et al. 1988; Wehrfritz et al. 1993; Jetten et al. 1997; Richardson et al. 1998). It thus can be suggested that the trace production of  $\text{NO}_2^-$  may be due to the rapid transformation of  $\text{NO}_2^-$  to  $\text{NO}_3^-$ , or due to the reduction of  $\text{NO}_2^-$  to gaseous nitrogen.

**Table 2**  $\text{NH}_4^+$ -N removal by *Proteus mirabilis* V7 in water samples

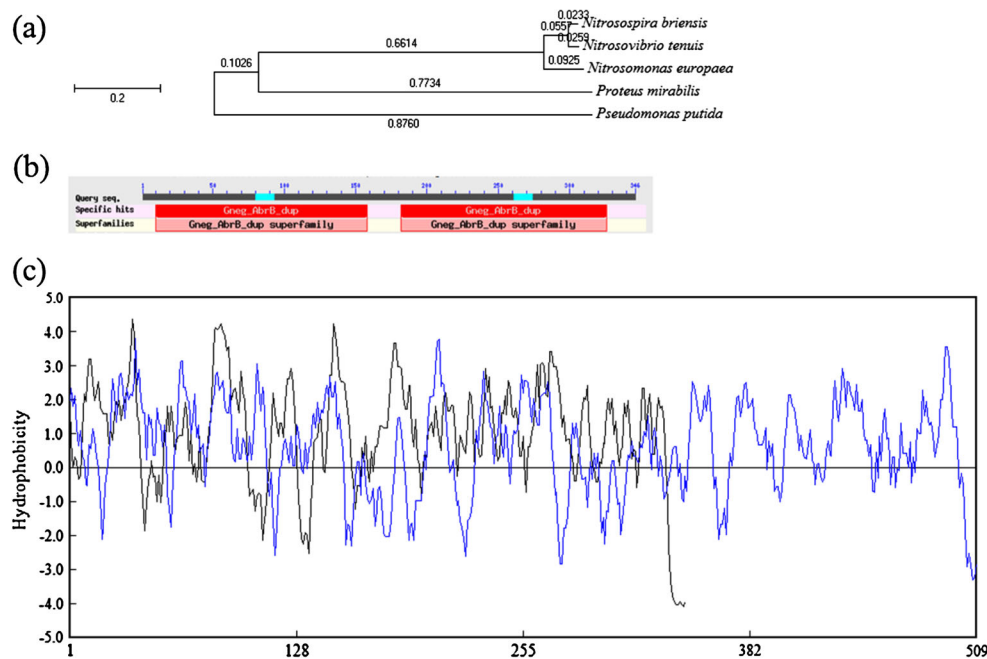
Sample ID	$\text{NH}_4^+$ -N ( $\text{mg L}^{-1}$ )		$\text{NO}_2^-$ -N ( $\text{mg L}^{-1}$ )		$\text{NO}_3^-$ -N ( $\text{mg L}^{-1}$ )	
	Initial	Final	Initial	Final	Initial	Final
1	64.40 $\pm$ 3.6	1.30 $\pm$ 0.06	0.35 $\pm$ 0.03	0.03 $\pm$ 0.002	30.80 $\pm$ 0.002	31.44 $\pm$ 3.4
2	18.00 $\pm$ 0.4	0.15 $\pm$ 0.01	0.00 $\pm$ 0.01	0.04 $\pm$ 0.009	0.74 $\pm$ 0.09	1.70 $\pm$ 0.3
3	9.82 $\pm$ 1.1	0.86 $\pm$ 0.13	0.00 $\pm$ 0.01	0.06 $\pm$ 0.006	0.00 $\pm$ 0.001	0.05 $\pm$ 0.003

#### Characterization of $\text{NH}_4^+$ -N removal by *Proteus mirabilis* strain V7

Removal of  $\text{NH}_4^+$ -N by *Proteus mirabilis* strain V7 grown in synthetic succinate medium at  $30^\circ\text{C}$  in the dark was investigated. As shown in Fig. 2, concentrations of  $\text{NH}_4^+$ -N in the supernatant of control medium at different selected time points remained almost the same for a 7-day observation, suggesting that  $\text{NH}_4^+$ -N was relatively steady under the conditions tested; however, inoculation of cells of *Proteus mirabilis* strain V7 caused a significant decrease in the level of  $\text{NH}_4^+$ -N with increasing time. After incubating for 4 days, only 25 % of the  $\text{NH}_4^+$ -N still remained in the culture. The removal efficiency under different initial concentrations of  $\text{NH}_4^+$ -N also showed significant differences. As shown in Fig. 3,  $\text{NH}_4^+$ -N at concentrations lower than  $100 \text{ mg L}^{-1}$  could be removed completely. Once the  $\text{NH}_4^+$ -N concentration rose to above  $200 \text{ mg L}^{-1}$ , the removal efficiency of *Proteus mirabilis* strain V7 decreased sharply. The removal efficiency of *Proteus mirabilis* strain V7 was below 20 % when the concentration of  $\text{NH}_4^+$ -N in the medium was higher than  $1,000 \text{ mg L}^{-1}$ . This result coincided with the fact that higher concentrations of  $\text{NH}_4^+$ -N also showed a negative effect on  $\text{NH}_4^+$ -N removal by *Bacillus cereus* strain I6 and *Acinetobacter calcoaceticus* strain STB1, i.e., the removal efficiency of *B. cereus* strain I6 dropped from 62 % to 24.8 % as concentrations of  $\text{NH}_4^+$ -N increased from 5 to  $25 \text{ mg L}^{-1}$ , and the removal efficiency of *A. calcoaceticus* strain STB1 also dropped from 93 % to 45 % as concentrations of  $\text{NH}_4^+$ -N increased from 100 to  $210 \text{ mg L}^{-1}$  (Hasan et al. 2012; Sarioglu et al. 2012).

#### $\text{NH}_4^+$ -N removal in seawater matrix and water samples

Removal of  $\text{NH}_4^+$ -N by *Proteus mirabilis* strain V7 grown in 2216E medium containing  $\text{NH}_4^+$ -N at concentrations of 20, 50, 100, 200, 500, 1000 and  $2000 \text{ mg L}^{-1}$ , was investigated. After culturing at  $30^\circ\text{C}$  for 64 h, supernatants were collected and analyzed for remaining  $\text{NH}_4^+$ -N. When the initial concentration of  $\text{NH}_4^+$ -N in 2216E medium was higher than  $50 \text{ mg L}^{-1}$ , hardly  $\text{NH}_4^+$ -N removal by *Proteus mirabilis* strain V7 was observed. Only  $20 \text{ mg L}^{-1}$   $\text{NH}_4^+$ -N in 2216E medium was removed completely by *Proteus mirabilis* strain V7 within 100 h (Fig. 4). Compared with  $\text{NH}_4^+$ -N removal in synthetic succinate medium, it could be concluded that



**Fig. 5** **a** Evolutionary relationship between AmoA<sub>Pm</sub> from *Proteus mirabilis* and AmoAs from other bacterial strains. Multiple sequence alignment was produced using CLUSTAL W2 (Larkin et al. 2007). The unrooted phylogenetic trees were constructed by the neighbour-joining method, based on the distances derived from the Dayhoff matrix, with MEGA 4.0 software (Tamura et al. 2007). The robustness of the branches was assessed by the bootstrap method with 1,000 replications. The scale bar represents a distance of 0.2 substitutions per site. The sequences have

the following NCBI accession numbers: *Nitrosomonas europaea* (AAA66194), *Nitrospira briensis* (AAB38709), *Nitrosobrio tenuis* (AAB38710) and *Pseudomonas putida* (CAA74724). **b** Schematic representation of the domains organized in the putative AmoA<sub>Pm</sub> from *Proteus mirabilis*. **c** The predicted hydropathy plots for AmoA from *Proteus mirabilis* (short black line) and AmoA from *Pseudomonas putida* (long blue line). Positive values indicate hydrophobic residues

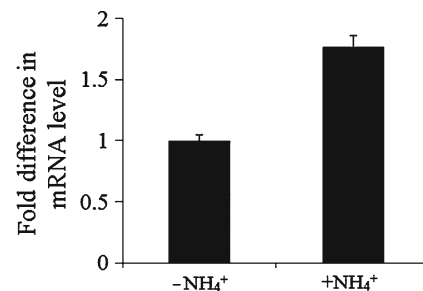
*Proteus mirabilis* strain V7 possessed a much lower NH<sub>4</sub><sup>+</sup>-N removal efficiency in the seawater matrix. According to a previous study by Wehrfritz et al. (1993), the coenzyme Q involved in NH<sub>4</sub><sup>+</sup> oxidization acquired an electron through oxidization of the supplied reducing organic carbon source during the heterotrophic nitrification process, and thus the organic carbon source exerted effects not only on the growth state but also on the nitrification activity of heterotrophic nitrifiers (Mével and Prieur 2000). Thus, the low nitrification efficiency of *Proteus mirabilis* strain V7 in the 2216E medium may be due to the oligotrophic characteristic of the medium.

Aquatic samples were collected from several estuaries around Yantai, and the concentrations of NH<sub>4</sub><sup>+</sup>-N, NO<sub>2</sub><sup>-</sup>-N and NO<sub>3</sub><sup>-</sup>-N were analyzed. Three water samples containing 64.40, 18.00, and 9.82 mg L<sup>-1</sup> NH<sub>4</sub><sup>+</sup>-N was chosen for the NH<sub>4</sub><sup>+</sup>-N removal experiment. Moreover, sample 1 also contained the highest concentrations of NO<sub>2</sub><sup>-</sup>-N and NO<sub>3</sub><sup>-</sup>-N. As shown in Table 2, almost all the NH<sub>4</sub><sup>+</sup>-N was removed within 48 h after inoculation with cells of *Proteus mirabilis* strain V7. In sample 1, which originally contained 0.35 mg L<sup>-1</sup> NO<sub>2</sub><sup>-</sup>-N, the concentrations of NO<sub>2</sub><sup>-</sup>-N were reduced 10-fold, while in samples 2 and 3, a trace increase in NO<sub>2</sub><sup>-</sup>-N was detected. Compared with the initial concentrations of NO<sub>3</sub><sup>-</sup>-N in the samples, a slight increase in the concentrations of NO<sub>3</sub><sup>-</sup>-N among all the three samples was observed, that might be

removed with longer incubated time. The increased production of NO<sub>2</sub><sup>-</sup>-N and NO<sub>3</sub><sup>-</sup>-N was postulated to be intermediates of NH<sub>4</sub><sup>+</sup> oxidization by *Proteus mirabilis* strain V7.

Analysis of the AmoA<sub>Pm</sub> sequence and its mRNA level in the presence of NH<sub>4</sub><sup>+</sup>

The key enzyme that catalyzes the first step of NH<sub>4</sub><sup>+</sup> oxidization is AMO (EC 1.14.99.39), which contains three subunits AmoA, AmoB and AmoC, with its active site located in AmoA (McTavish et al. 1993; Klotz et al. 1997; Hirota et al. 2000). *amoA*<sub>Pm</sub>, encoding a putative AmoA from *Proteus mirabilis*, was cloned and submitted for DNA sequencing.



**Fig. 6** mRNA analysis of the putative *amoA*<sub>Pm</sub> gene. The mRNA level of the putative *amoA*<sub>Pm</sub> was normalized to that of 16S rRNA and all data are presented in terms of relative mRNA. Data are the means for three independent experiments and are presented as the means±SE

Analysis of the nucleotide sequence obtained revealed it was an open reading frame (ORF) with a length of 1,041 bp. The encoded protein had an estimated molecular mass of approximately 38 kDa and pI of 9.96, and shared 99 % overall sequence similarity with the membrane-associated ammonia monooxygenases from three *Proteus mirabilis* strains with accession numbers WP\_004245042, YP\_008396899 and YP\_002149941, respectively. A phylogenetic analysis of AmoA<sub>Pm</sub> using the other known AmoAs is shown in Fig. 5a. Compared with AmoA from *Pseudomonas putida*, the putative AmoA from *Proteus mirabilis* showed a higher homology to the AmoAs from autotrophic ammonia oxidizers. The deduced amino acid sequence of the identified ORF contained two regions that belonged to Gneg\_AbrB\_dup superfamily, which showed distinct sequence similarities to the corresponding motif of AmoAs from autotrophic ammonia oxidizers (Fig. 5b). In addition, the hydrophobicity profile of the AmoA<sub>Pm</sub> was similar to the corresponding regions of AmoA from *Pseudomonas putida*, which was also similar to the AmoAs from *Nitrosomonas europaea*, *Nitrospira briensis* and *Nitrosovibrio tenuis* (Daum et al. 1998) (Fig. 5c). After being induced by IPTG, no recombinant protein was detected in BL21(DE3)/pETAmoA<sub>Pm</sub> after SDS-PAGE analysis. This phenomenon was consistent with the fact that the expression of *amo* genes was toxic to the *E. coli* strain, thus none of the corresponding gene fragments has been cloned and heterogeneously expressed successfully in *E. coli* combined with the pET system (Bergmann and Hooper 1994; Gupta and Khare 2006). That may account for why only two AmoAs from heterotrophic nitrifiers, i.e., the AmoA from *Pseudomonas putida* and *Paracoccus denitrificans*, have been identified, and AmoAs from heterotrophic nitrifying bacteria exhibited significant differences in strain specificity (Huertas and Duque 1998; Sardesai and Bhosle 2002; Gupta and Khare 2006). In order to investigate whether the transcription of *amoA*<sub>Pm</sub> was affected by NH<sub>4</sub><sup>+</sup>, *Proteus mirabilis* strain V7 was grown in phosphate salt medium and synthetic succinate medium at 30 °C for 12 h, respectively. Total RNA was extracted and used for real-time RT-PCR to analyze the relative mRNA level of *amoA*<sub>Pm</sub>. The result showed that a 1.8-fold increase in the mRNA level of *amoA*<sub>Pm</sub> was observed in the presence of NH<sub>4</sub><sup>+</sup> (Fig. 6). This result was coincided with the fact that the mRNA levels of *amo* and enzyme activity were positively correlated with concentration of NH<sub>4</sub><sup>+</sup> rose up to 1 mM after incubation for 3 h (Sayavedra-Soto et al. 1996)

## Conclusions

In this study, a heterotrophic nitrifying bacterium *Proteus mirabilis* strain V7 was identified and characterized. *Proteus mirabilis* strain V7 removed 100 % of NH<sub>4</sub><sup>+</sup>-N at concentrations below 100 mg L<sup>-1</sup> and 51 % of TIN in the culture

medium within 96 h, which indicated that *Proteus mirabilis* strain V7 was a heterotrophic nitrifying bacterium that may possess the ability of both heterotrophic nitrification and aerobic denitrification. *Proteus mirabilis* strain V7 could also remove NH<sub>4</sub><sup>+</sup>-N in seawater matrix but with lower removal efficiency. In water samples contaminated by NH<sub>4</sub><sup>+</sup>-N, *Proteus mirabilis* strain V7 could remove all the NH<sub>4</sub><sup>+</sup>-N with a small increase in the levels of NO<sub>2</sub><sup>-</sup>-N and NO<sub>3</sub><sup>-</sup>-N. The mRNA level of *amoA*<sub>Pm</sub> was increased in the presence of NH<sub>4</sub><sup>+</sup>, which indicated that AmoA<sub>Pm</sub> from *Proteus mirabilis* strain V7 may be involved in NH<sub>4</sub><sup>+</sup> oxidization.

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## References

- Alcaraz G, Espinoza V, Vanegas C, Carrara XC (1999) Acute effect of ammonia and nitrite on respiration of *Penaeus setiferus* postlarvae under different oxygen levels. *J World Aquacult Soc* 30:98–106
- Alexander M (1977) Nitrification. In: *Soil microbiology*, 2nd edn. Wiley, New York, pp 251–271
- Andrade do Canto CS, Rodrigues JA, Ratusznei SM, Zaiat M, Foresti E (2008) Feasibility of nitrification/denitrification in a sequencing batch biofilm reactor with liquid circulation applied to post-treatment. *Bioresour Technol* 99:644–654
- APHA (1995) Standard methods for the examination of water and wastewater, 19th edn. American Public Health Association, AWWA, WEF, Washington, DC
- Bergmann DG, Hooper AB (1994) Sequence of the gene, *amo B*, for the 43-kDa polypeptide of ammonia monooxygenase of *Nitrosomonas europaea*. *Biochem Biophys Res Commun* 204:759–762
- Castignetti D, Hollocher T (1982) Nitrogen redox metabolism of a heterotrophic, nitrifying-denitrifying *Alcaligenes* sp. from soil. *Appl Environ Microbiol* 44:923–928
- Castignetti D, Hollocher T (1984) Heterotrophic nitrification among denitrifiers. *Appl Environ Microbiol* 47:620–623
- Daum M, Zimmer W, Papen H, Kloos K, Nawrath K, Bothe H (1998) Physiological and molecular biological characterization of ammonia oxidation of the heterotrophic nitrifier *Pseudomonas putida*. *Curr Microbiol* 37:281–288
- Gupta A, Khare SK (2006) A protease stable in organic solvents from solvent tolerant strain of *Pseudomonas aeruginosa*. *Bioresour Technol* 97:1788–1793
- Hasan HA, Sheikh Abdullah SR, Kofli NT, Kamarudin SK (2012) Removal of ion in drinking water treatment using locally isolated heterotrophic nitrifier. *Desalin Water Treat* 50:294–301
- Hirota R, Yamagata A, Kato J, Kuroda A, Ikeda T, Takiguchi N, Ohtake H (2000) Physical map location of the multicopy genes coding for ammonia monooxygenase and hydroxylamine oxidoreductase in the ammonia-oxidizing bacterium *Nitrosomonas* sp. strain ENI-11. *J Bacteriol* 182:825–828
- Huertas M, Duque E (1998) Survival in soil of different toluene-degrading *Pseudomonas* strains after solvent shock. *Appl Environ Microbiol* 64:38–42
- Jetten MS, Logemann S, Muyzer G, Robertson LA, de Vries S, van

- Loosdrecht MC, Kuenen JG (1997) Novel principles in the microbial conversion of nitrogen compounds. *Antonie Van Leeuwenhoek* 71:75–93
- Jorgensen TC, Weatherley LR (2003) Ammonia removal from wastewater by ion exchange in the presence of organic contaminants. *Water Res* 37:1723–1728
- Khardenavis AA, Kapley A, Purohit KJ (2007) Simultaneous nitrification and denitrification by diverse *Diaphorobacter* sp. *Appl Microbiol Biotechnol* 77:403–409
- Kim JK, Park KJ, Cho KS, Nam SW, Park TJ, Bajpai R (2005) Aerobic nitrification–denitrification by heterotrophic *Bacillus* strains. *Bioresour Technol* 96:1897–1906
- Klotz MG, Alzerreca J, Norton JM (1997) A gene encoding a membrane protein exists upstream of the amoA/amoB genes in ammonia oxidising bacteria: a third member of the amo operon? *FEMS Microbiol Lett* 150:65–73
- Larkin MA, Blackshields G, Brown NP, Chenna R, McGettigan PA, McWilliam H, Valentin F, Wallace IM, Wilm A, Lopez R, Thompson JD, Gibson TJ, Higgins DG (2007) Clustal W and Clustal X version 2.0. *Bioinformatics* 23:2947–2948
- McTavish H, Fuchs JA, Hooper AB (1993) Sequence of the gene coding for ammonia monooxygenase in *Nitrosomonas europaea*. *J Bacteriol* 175:2436–2444
- Mevel G, Prieur D (2000) Heterotrophic nitrification by a thermophilic *Bacillus* species as influenced by different culture conditions. *Can J Microbiol* 46:465–473
- Otte S, Grobbsen NG, Robertson LA, Jetten MSM, Kuenen JG (1996) Nitrous oxide production by *Alcaligenes faecalis* under transient and dynamic aerobic and anaerobic conditions. *Appl Environ Microbiol* 62:2421–2426
- Papen H, von Berg R (1998) A most probable number method for the estimation of cell numbers of heterotrophic nitrifying bacteria in soil. *Plant Soil* 799:723–730
- Papen H, von Berg R, Hinkel I, Thoene B, Rennenberg H (1989) Heterotrophic nitrification by *Alcaligenes faecalis*: NO<sub>2</sub><sup>-</sup>, NO<sub>3</sub><sup>-</sup>, N<sub>2</sub>O, and NO production in exponentially growing cultures. *Appl Environ Microbiol* 55:2068–2072
- Patureau D, Bernet N, Moletta R (1997) Combined nitrification and denitrification in a single aerated reactor using the aerobic denitrifier *Comamonas* sp. strain SGLY2. *Water Res* 31:1363–1370
- Richardson DJ, Wehrfritz JM, Keech A, Crossman LC, Roldan MD, Sears HJ, Butler CS, Reilly A, Moir JWB, Berks BC, Ferguson SJ, Thomson AJ, Spiro S (1998) The diversity of redox proteins involved in bacterial heterotrophic nitrification and aerobic denitrification. *Biochem Soc Trans* 26:401–408
- Robertson LA, Gijs Kuenen J (1984) Aerobic denitrification: a controversy revived. *Arch Microbiol* 139:351–354
- Robertson LA, Van Neil EWJ, Torremans RAM, Kuenen JG (1988) Simultaneous nitrification and denitrification in aerobic chemostat cultures of *Thiosphaera pantotropha*. *Appl Environ Microbiol* 54:2812–2818
- Rostron WM, Stuckey DC, Young AA (2001) Nitrification of high strength ammonia wastewaters: comparative study of immobilization media. *Water Res* 35:1169–1178
- Sardessai Y, Bhosle S (2002) Organic solvent-tolerant bacteria in mangrove ecosystem. *Curr Sci* 82:622–623
- Sarioglu OF, Suluyayla R, Tekinay T (2012) Heterotrophic ammonium removal by a novel hatchery isolate *Acinetobacter calcoaceticus* STB1. *Int Biodeter Biodegrad* 71:67–71
- Sayavedra-Soto LA, Hommes NG, Russell SA, Arp DJ (1996) Induction of ammonia monooxygenase and hydroxylamine oxidoreductase mRNAs by ammonium in *Nitrosomonas europaea*. *Mol Microbiol* 20:541–548
- Schaechter M (2009) *Encyclopedia of microbiology*, 3rd edn. Elsevier, Amsterdam
- Su JJ, Liu BY, Liu CY (2001) Comparison of aerobic denitrification under high oxygen atmosphere by *Thiosphaera pantotropha* ATCC 35512 and *Pseudomonas stutzeri* SU2 newly isolated from the activated sludge of a piggery wastewater treatment system. *J Appl Microbiol* 90:457–462
- Tamura K, Dudley J, Nei M, Kumar S (2007) MEGA4: molecular evolutionary genetics analysis (MEGA) software version 4.0. *Mol Biol Evol* 24:1596–1599
- Taylor SM, He Y, Zhao B, Huang J (2009) Heterotrophic ammonium removal characteristics of an aerobic heterotrophic nitrifying denitrifying bacterium, *Providencia rettgeri* YL. *J Environ Sci* 21:1336–1341
- US EPA (1993) *Process design manual for nitrogen control*. United States Environmental Protection Agency, Washington, DC
- US EPA (2009) *Draft 2009 Update of aquatic life ambient water quality criteria for ammonia in freshwater*. United States Environmental Protection Agency, Washington, DC
- Ward BB (1996) Nitrification and ammonification in aquatic systems. *Life Support Biosphere Sci: Int J Earth Space* 3:25–29
- Wehrfritz JM, Reilly A, Spiro S, Richardson DJ (1993) Purification of hydroxylamine oxidase from *Thiosphaera pantotropha*, identification of electron acceptors that couple heterotrophic nitrification to aerobic denitrification. *FEBS Lett* 335:246–250
- Whitfield M (1974) The hydrolysis of ammonium ions in sea water—a theoretical study. *J Mar Biol Assoc UK* 54:565–580
- WHO (1990) *Ammonia Health and Safety Guide*. World Health Organization, Geneva. online at: <http://www.inchem.org/documents/hsg/hsg/hsg037.htm#SubSectionNumber:1.3.2>
- Winogradsky S (1890) Recherches sur les organismes de la nitrification. *Ann Inst Pasteur (Paris)* 4:213–231
- Witzel KP, Overbeck HJ (1979) Heterotrophic nitrification by *Arthrobacter* sp. (strain 9006) as influenced by different cultural conditions, growth state and acetate metabolism. *Arch Microbiol* 122:137–114
- Yang XP, Wang SM, Zhang DW, Zhou LX (2011) Isolation and nitrogen removal characteristics of an aerobic heterotrophic nitrifying-denitrifying bacterium, *Bacillus subtilis* A1. *Bioresour Technol* 102:854–862
- Zhang WW, Sun L (2007) Cloning, characterization, and molecular application of a beta-agarase gene from *Vibrio* sp. strain V134. *Appl Environ Microbiol* 73:2825–2831
- Zhang WW, Sun K, Cheng S, Sun L (2008) Characterization of DegQVh, a serine protease and a protective immunogen from a pathogenic *Vibrio harveyi* strain. *Appl Environ Microbiol* 74:6254–6262
- Zhao B, He YL, Hughes J, Zhang XF (2010) Heterotrophic nitrogen removal by a newly isolated *Acinetobacter calcoaceticus* HNR. *Bioresour Technol* 101:5194–5200